

Mining waste acting as a precursor of environmental stress in sediments

Rejeitos de mineração atuando como precursores de estresse ambiental em sedimentos

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ABSTRACT

Waste generated by mineral extraction is globally associated with environmental disturbances due to its deleterious effect on water resources. However, research focused on the influence of mine tailings resulting from the extraction of semi-precious stones on fluvial systems is still incipient in the environmental literature. From this perspective, this study quantified the average concentrations of major oxides present in the fine fractions of the sediment samples from the Várzea river, in the State of Rio Grande do Sul, southern Brazil, using wavelength dispersive X-ray fluorescence spectrometry. This region is acknowledged as the largest rock amethyst mining area in the world. Additionally, geochemical indices were established to characterize potential sources of production, maturity, degree of weathering, and sediment pollution. To evaluate the influence of mine tailings on the Várzea river sediments, the contents of Al_2O_3 , Fe_2O_3 , MnO, P_2O_5 , CaO, SiO_2 , K_2O , CuO, ZnO, and TiO_2 major oxides present in sediment samples were determined and compared to the local background values; the values varied significantly ($p < 0.05$), classifying them as polluted and medium polluted. Also, the sediment samples with evident characteristics of extreme chemical weathering consist mainly of clay minerals and mafic igneous rocks, and similarities were found between sediment samples and tailings from the mineral extraction zone. The Principal Component Analysis and the cluster analysis also suggest the existence of three distinct mineral oxide groups, differentiating the zones leaving and upstream the mining zone from the other sampling points.

Keywords: mineral prospecting; Ametista do Sul; geochemistry; WD-XRF.

RESUMO

Os rejeitos de atividades de extração mineral são mundialmente associados a distúrbios ambientais devido a sua capacidade de atuação deletéria sobre os recursos hídricos. No entanto, o número de estudos sobre a influência dos rejeitos de zonas de extração de rochas semipreciosas sobre sistemas flúvicos ainda é incipiente na literatura da área ambiental. Nessa perspectiva, foram quantificadas pela técnica de espectrometria de fluorescência de raios-X por dispersão de comprimento de onda as concentrações médias de óxidos minerais majoritários presentes nas frações finas das amostras de sedimentos do Rio da Várzea, Sul do Brasil, maior região mundial extratora de rochas ametista. Somado a isso, avaliou-se a qualidade ambiental e caracterizaram-se as potenciais fontes por meio de índices geoquímicos. Os resultados consideram a provável influência dos rejeitos de mineração sobre os sedimentos do rio da Várzea, classificando-os como não poluído a moderadamente poluído e, quando comparados ao *background* local, apresentam variâncias significativas ($p < 0,05$) para os óxidos minerais Al_2O_3 , MnO, CaO, SiO_2 , K_2O e TiO_2 . Além disso, as amostras de sedimentos que apresentaram características de intemperismo químico extremo são formadas por argilominerais e rochas ígneas máficas, e foram encontradas similaridades entre as amostras de sedimentos e os rejeitos da zona de extração mineral. Conclui-se que os rejeitos de rochas basálticas, acúmulos de frente aos garimpos, funcionam como fonte de estresse ambiental no sistema hídrico.

Palavras-chave: prospecção mineral; Ametista do Sul; geoquímica; WD-XRF.

INTRODUCTION

Mineral extraction activities affect the environmental quality of adjacent water systems, causing deleterious changes to ecosystems (BECK *et al.*, 2020; GLOAGUEN; MOTTA; COUTO, 2021; KUSIN *et al.*, 2019; SILVA *et al.*, 2019). In these activities,

the environmental contamination results from the rainwater ability to fractionate, solubilize and transport minerals and inorganic compounds (metals, semi-metals, and non-metals) found in the tailings from mining (AMOSOVA *et al.*, 2019; BESSA *et al.*, 2020; GUILHEN *et al.*, 2019; SILVA *et al.*, 2019).

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Mine tailings are associated with environmental disturbances, such as acid mine drainage, reduced hydrogen and redox potential, restricted oxygen concentrations, elevated salinization, and changed alkalinity due to reactions with carbonates while also favoring sediment production/loading (SANG *et al.*, 2018; SILVA *et al.*, 2019; ZHAO *et al.*, 2020). Mine tailings carried into the aquatic environment, via surface processes (adsorption, complexation, and reprecipitation), tend to decant and become part of the fluvic sediments (BECK *et al.*, 2020; REMOR *et al.*, 2018; SUNDARARAJAN *et al.*, 2017). In Brazil, the contamination that occurred in recent years in Brumadinho, Mariana, and Santo Antônio do Grama (cities in the state of Minas Gerais) and Barcarena (city in the state of Pará) due to mine tailings demonstrate the difficulty of addressing this environmental problem (ALVES; FERREIRA; ARAÚJO, 2021; PASTRAN; MALLETT 2020; SILVA *et al.*, 2019).

Sediments are complex, multi-element and polygranular environmental matrices, whose ability to store and diffuse inorganic compounds turn them into environmental stressors that can directly interfere with the aquatic ecosystem dynamics (HERATH *et al.*, 2018; SILVA *et al.*, 2016; SILVA *et al.*, 2019). The term 'environmental stressor' describes chemical elements, biological or physical entities in sediments that may cause adverse effects (BURTON; JOHNSTON, 2010; SILVA *et al.*, 2016; USEPA, 1992).

Nevertheless, sediments are also an essential, integral, and dynamic part of water systems, such as rivers and lakes, while human activities, such as mineral extraction, directly affect sediment quantity and composition, making environmental planning and management actions necessary (LONE *et al.*, 2018; MIZAEAL *et al.*, 2020; SILVA *et al.*, 2016). In watersheds with mining activities, the concentrations of certain chemical compounds can indicate tailings influence as a residue/waste source, contributing to the water system and interfering in the environmental quality (GOMES; ALMEIDA; SPERANDIO, 2018; KUSIN *et al.*, 2019; SILVA *et al.*, 2019).

Understanding the mobility of major oxides during the weathering processes has been proposed as an alternative to identify sediment formation zones (GOMES; ALMEIDA; SPERANDIO, 2018; SANG *et al.*, 2018; SILVA *et al.*, 2019; SOUSA *et al.*, 2018). Additionally, sediment quality is often assessed by environmental geochemical indices based on average concentrations of local inorganic chemical compounds, the background values, compared to bottom levels as a tool for water resource management (BILLAH; KOKUSHI; UNO, 2019; MIZAEAL *et al.*, 2020).

This method considers that variations of the average concentrations of fluvial sediments are associated with adverse effects on aquatic ecosystems, allowing to track down where such environmental stress sources are located (HERATH *et al.*, 2018; KUSIN *et al.*, 2019; SUNDARARAJAN *et al.*, 2017). For this purpose, this study uses this approach in the semi-precious rock extraction zones to assess the influence of mine tailings on local water systems since there are few studies in the literature applying this methodology (DIAS *et al.*, 2019; HARTMANN; PERTILLE; DUARTE, 2017; SILVA *et al.*, 2019). Besides, geochemical concentrations in sediments were characterized and quantified to determine the potential influence of mining tailings, the respective sources, and to assess the quality of sediments by determining the geochemical indices, on the Várzea river, a water system located in southern Brazil, the region with the largest amethyst extraction in the world (DIAS *et al.*, 2019; SILVA *et al.*, 2019; VOSOOGH; SAEEDI; LAK, 2016).

Although Silva *et al.* (2019) investigated sediment contamination in the mining region of the Várzea river, using an energy dispersive X-ray fluorescence spectrometer (EDXRF — Shimadzu equipment, model EDX-720, Rannye Series), and established environmental quality indices, these authors did not address the geochemical characterization through indices based on the mean concentrations of the major mineral oxides that structure sediment samples. This gap is intended to be filled in this upcoming study, through instrumental analyzes performed with a wavelength dispersive X-ray fluorescence spectrometer (WDXRF — BRUKER, model S8 TIGER), allowing to determine the condition of weathering and identifying its source and main characteristics.

MATERIAL AND METHODS

Study site

Located in the northern region of Rio Grande do Sul (Southern Brazil), the Várzea river length is approximately 165 km and it drains an area of 9,324 km² (Figure 1). It is one of the main left-bank tributaries of the Uruguay river, in the hydrographic region of Uruguay (FEPAM, 2020; SEMA, 2012; SILVA *et al.*, 2019).

The Várzea river flows through regions with a dense red dystrophic red latosol soil, characterized by strong acidity and low nutrient reserve that usually requires fertility corrections. The main agricultural activities are crops

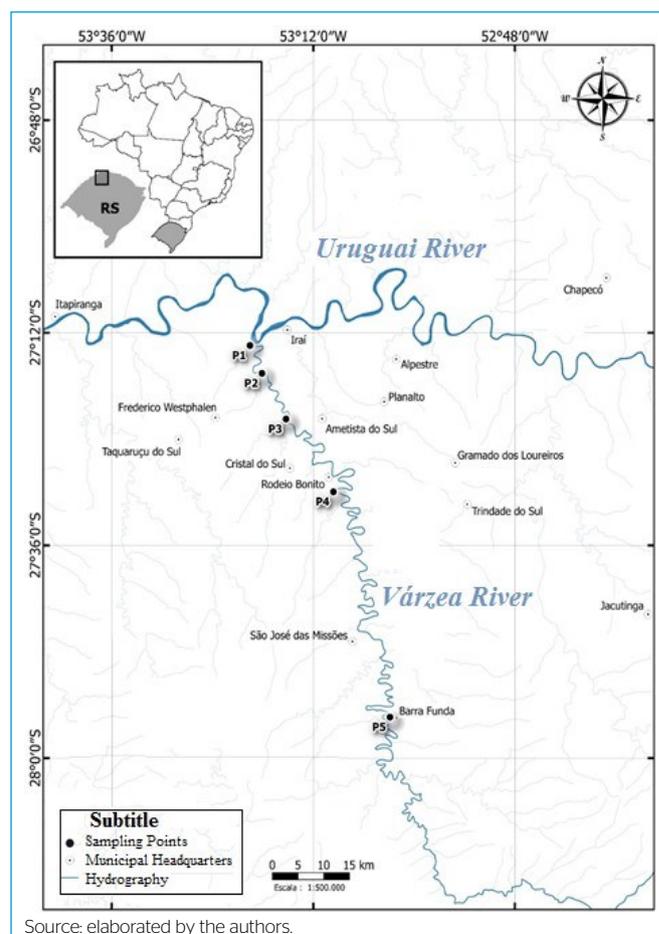


Figure 1 – Hydrographic map containing the spatial and fluvial distribution of the Várzea river showing the sediment sampling points.

(wheat, corn, and soybeans) and animal husbandry (chicken, pork, and cattle) (SANTOS *et al.*, 2018; MSRS, 2020). In the northern portion of the Várzea river, semi-precious rocks (amethyst, quartz, agate, and calcite) are mined in horizontal underground mines. Figure 1 shows the municipalities of Ametista do Sul, the world's largest producer of amethyst, as well as Frederico Westphalen, Rodeio Bonito, Cristal do Sul, Planalto, Iraí, Trindade do Sul, and Gramado dos Loureiros (HARTMANN; PERTILLE; DUARTE, 2017; KORCHAGIN; CANER; BORTOLUZZI, 2019).

The genesis of the semi-precious rocks is attributed to volcanic deposits in the Paraná basin from the basaltic flows related to the Serra Geral formation, resulting from the interaction of hydrothermal events associated with the Guarani aquifer (BAGGIO *et al.*, 2015; DIAS *et al.*, 2019; KORCHAGIN; CANER; BORTOLUZZI, 2019). Figure 2 showed the semi-precious rocks extracted in the region.

Several studies characterize the concentrations of elements present in the basalt residues to determine the formation and geochemical extent of the semi-precious rocks (HARTMANN *et al.*, 2015; KORCHAGIN; CANER; BORTOLUZZI, 2019; SILVA *et al.*, 2019). Table 1 shows the average concentration of compounds present in the basalt waste.

However, few studies so far have been able to demonstrate the influence of mineral extraction activities on the quality of sediment in water systems, and also address the influence of this activity as environmental liability generated from tons of tailings from basalt rocks (rocks of volcanic origin rich in iron and magnesium silicates) stored/exposed to weather in front of mines (Figure 3).

Under favorable environmental conditions, these tailings tend to fragment into small particles that end up carried into the Várzea river, possibly leading to harmful environmental effects (BAGGIO *et al.*, 2015; KORCHAGIN; CANER; BORTOLUZZI, 2019; SILVA *et al.*, 2019).

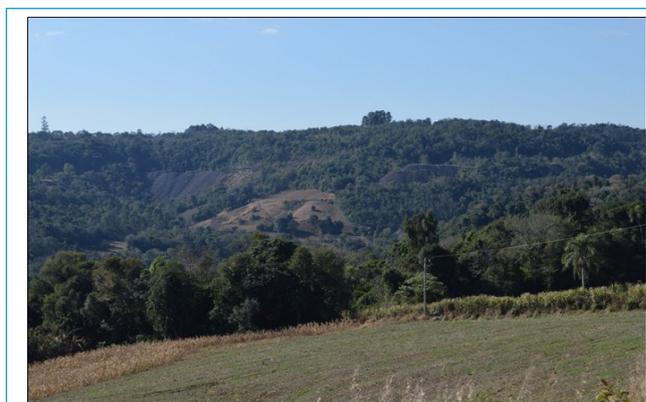
Sediment sampling, storing, and preparation

Figure 1 shows the Várzea river watershed, highlighting the sampling points deemed relevant for this study (upstream, downstream, and inside the mineral extraction zones) regarding both spatial variability and ease of access. The sediment sampling processes followed clean technique protocols, using a stainless-steel Petersen dredge (3 kg) (BRANDÃO *et al.*, 2011; SILVA *et al.*, 2019).

Table 1 - Average concentrations of major oxides present in samples of basalt rock tailings from the mining zone in the northern portion of the Várzea river.

	Baggio <i>et al.</i> (2015)	Hartmann <i>et al.</i> (2015)	Korchagin; Caner; Bortoluzzi (2019)
	Amethyst: Mina Museum ($\bar{X} \pm CI$)	Amethyst: Mina Museum ($\bar{X} \pm CI$)	Amethyst: Geodes ($\bar{X} \pm CI$)
Oxides (%)			
Na ₂ O	2.36 ± 0.04	2.21 ± 0.03	3.67 ± 0.18
MgO	5.56 ± 0.17	4.41 ± 0.36	5.55 ± 0.36
Al ₂ O ₃	13.08 ± 0.18	14.99 ± 0.54	15.90 ± 0.52
SiO ₂	49.42 ± 0.46	46.63 ± 1.28	47.42 ± 1.26
P ₂ O ₅	0.28 ± 0.01	3.58 ± 0.06	0.87 ± 0.02
K ₂ O	1.13 ± 0.19	1.19 ± 0.23	1.47 ± 0.18
CaO	9.57 ± 0.07	8.10 ± 0.18	8.27 ± 0.28
TiO ₂	2.28 ± 0.06	12.21 ± 0.27	3.54 ± 0.17
MnO	0.20 ± 0.02	0.56 ± 0.02	0.20 ± 0.02
Fe ₂ O ₃	14.86 ± 0.61	0.20 ± 0.03	13.85 ± 0.87
ZnO	ND	ND	0.05 ± 0.00
SrO	ND	ND	0.06 ± 0.01

\bar{X} : average concentration; CI: confidence interval; ND: non-determined. Source: reference values for basalt tailings, average concentrations calculated for three rock samples, Baggio *et al.* (2015); 16 samples, Hartmann *et al.* (2015); five samples, Korchagin; Caner; Bortoluzzi (2019).



Source: elaborated by the authors.

Figure 3 - Mine tailings accumulated in front of the mine in Ametista do Sul, Southern Brazil.



Source: elaborated by the authors.

Figure 2 - The semi-precise rocks extracted in the Várzea river mining region (rough rocks and cut jewels), Southern Brazil.

The sampling points in Figure 1 correspond to the municipalities of Iraí, on the BR-386 bridge (at the Iraí and Frederico Westphalen border), downstream from the mining zone (P1); Iraí on the Estrada da Ponte Velha (at the Ametista do Sul and Iraí border), exiting the mining zone (P2); the main mining town of Ametista do Sul (P3); Rodeio Bonito and Cristal do Sul, at the beginning of the mining zone (P4); and Barra Funda, upstream from the mining zone (P5). The sampling dates and geographic location are shown in Table 2.

The sampling protocol consisted of sampling in triple throw, perpendicular to the river cross-section (margins and center), generating composite samples per sampling point that were stored in 2 L polypropylene Ziploc bags, transported in cooling boxes to the laboratory, where they were kept refrigerated in a freezer (-25°C) until the analysis (SILVA *et al.*, 2016; SILVA *et al.*, 2019). For the analysis, the sediment samples were dried in a circulation oven, manually disintegrated using an agate mortar and pestle, and separated into fine fractions (< 0.063 mm) in a sieve shaker system.

Characterizing and quantifying major oxides in sediment samples

The major oxides present in the fine fractions of the sediment samples were characterized using a BRUKER wavelength dispersive X-ray fluorescence spectrometer (WD-XRF), model S8 TIGER (1 KW). The system was equipped for three ranges, the Na and Mg (XS 55 crystal, P10 gas detector), Al to S (PET crystal, P10 gas detector) and above K (LiF200 crystal, scintillation detector), as well as mask in the X-ray tube to irradiate a 28 mm diameter area of the sample. Measurements were performed on powdered sediment samples (4 g, < 63 µm fractions) packed in specific containers covered with 2.5 µm thick mylar film, under He atmosphere, in the Quant Express routine.

The analytical method was calibrated to quantify the major oxides (Al₂O₃, SiO₂, P₂O₅, K₂O, CaO, TiO₂, Fe₂O₃, MgO, MnO, ZnO, and SrO) found in the sediment samples, compared to the certified reference materials (CRM), Mineral Barium Carbonate (IAEA-Co9) of the International Atomic Energy Agency (IAEA); Sediment Green River (SGR-1b) and Sediment Cody (SCo-1) shales, both from the United States Geological Survey (USGS).

The WD-XRF is often used for the geochemical characterization of environmental matrices (sediments, soils, rocks) due to its high precision since the technique does not interfere with the compounds' chemical bonds, besides the low analytical cost compared to other techniques (CROUDACE *et al.*, 2019; GUILHEN *et al.*, 2019; OYEDOTUN, 2018). A total of 45 WD-XRF analyses were performed on the fine fractions of 15 composite sediment samples from the Várzea river, and the results were accepted as consistent according to the level of confidence for environmental matrices (INMETRO, 2018; SILVA *et al.*, 2019).

The quality of the analytical process was linked to the instrumental recovery of major oxides in the CRM (IAEA-Co9, SGR-1b, SCo-1) (Supplementary Figure 1). A good agreement is observed between CRM values and WD-XRF analyses in the laboratory (between 80 and 120%) for Na₂O, MgO, Al₂O₃, SiO₂, P₂O₅, K₂O, CaO, TiO₂, MnO, Fe₂O₃, CuO, ZnO, and SrO.

Statistical analysis and graphs

The outliers were removed by applying the Grubbs test to the datasets of major oxides in the QuickCalcs Outlier Calculator system (DOTMATICS, 2020). The SciDavis software was used for the plots (SOURCEFORGE, 2018), while other data were processed using Microsoft Office softwares. The datasets were submitted to the Jarque-Bera Normality test, Analysis of Variance (ANOVA), and Tukey test using the PAleontological STatistics (PAST) statistical software, Version 3.08 (HAMMER, 2015).

The dendrogram of the Hierarchical Cluster Analysis (Manhattan metric and Ward method) and the Principal Component Analysis (PCA) were performed with a 15 x 13 matrix, referring to 15 samples (sampling points in the sampling campaigns 1, 2, and 3) and 13 variables (each chemical component concentration). The data were self-scaled, and the analyses were performed using the R software (THE R FOUNDATION, [s.d.]).

Environmental geochemical indexes

The Chemical Index of Alteration (CIA) determines the degree of chemical weathering of river sediments as it reveals the resistance of aluminum silicates to the formation of clay minerals (GOMES; ALMEIDA; SPERANDIO, 2018; KUSIN *et al.*, 2019; SANG *et al.*, 2018). Equation 1 is used to calculate the CIA scores.

$$CIA = \left(\frac{Al_2O_3}{Al_2O_3 + CaO + Na_2O + K_2O} \right) \times 100 \quad (1)$$

The CIA scores are linked with the concentration of major oxides in the fluvial sediments and allow a better understanding of the sediment formation zones (Table 3).

Also, chemical weathering is usually assessed by equations that establish elementary ratios (ER), as in Equations 2 and 3.

$$ER_1 = \left(\frac{K_2O}{Al_2O_3} \right) \quad (2)$$

$$ER_2 = \left(\frac{Al_2O_3}{TiO_2} \right) \quad (3)$$

Both ER₁ and ER₂ (Table 3) are based on the concentration of major oxides and establish composition indexes that characterize the sediment source rock (GOMES; ALMEIDA; SPERANDIO, 2018; KUSIN *et al.*, 2019; SHEN *et al.*, 2018). The ER₃ and ER₄ ratios, as in Equations 4 and 5, together with the CaO

Table 2 - Sampling campaigns, geographical coordinates of the sampling points of the Várzea river sediments.

Sampling points	Code	Coordinates		Campaigns	Date	Material
Barra Funda	BF	27°55'22" S	53°02'52" W	1°	02/07/2016	Sediments
Rodeio Bonito	RB	27°26'33" S	53°10'13" W	2°	17/09/2016	Sediments
Ametista do Sul	A	27°21'44" S	53°15'16" W			
Iraí – Ponte Velha	IPV	27°16'33" S	53°18'07" W	3°	01/11/2016	Sediments
Iraí – BR-386	IBR	27°13'25" S	53°19'31" W			

Source: elaborated by the authors.

Table 3 - Scores for environmental geochemical indexes.

Value	Chemical Index of Alteration (CIA)	Value	Elementary ratio (ER ₁)	Value	Elementary ratio (ER ₂)
50 ≤	No chemical weathering	0.0 - 0.3	Clay minerals	3-8	Originated from mafic igneous rocks
50 - 60	Low chemical weathering			8-21	Intermediate rocks
60 - 80	Moderate chemical weathering	0.3 - 0.9	Feldspars	21-70	Felsic igneous rocks
80 >	Extreme chemical weathering				

Source: elaborated by the authors.

content, have been used to assess the maturity and chemical composition of sediments (SANG *et al.*, 2018; ZANARDO *et al.*, 2017).

$$ER_3 = \left(\frac{SiO_2}{Al_2O_3} \right) \quad (4)$$

$$ER_4 = \left(\frac{Fe_2O_3}{K_2O} \right) \quad (5)$$

Sediments from different sources have characteristic geochemical signatures (BESSA *et al.*, 2020; GLOAGUEN; MOTTA; COUTO, 2021; SHEN *et al.*, 2018; SILVA; BATEZELLI; LADEIRA, 2015; SILVA *et al.*, 2019). This aspect can be explored to define regions of environmental similarity in diagrams of the K₂O/Na₂O versus SiO₂ type (ARISEKAR *et al.*, 2021; SILVA; BATEZELLI; LADEIRA, 2015; ZANARDO *et al.*, 2017).

Contamination in sediment samples from the Várzea river was evaluated in relation to the mean concentration of major oxides using the Nemerow Pollution Index (NPI) (ARISEKAR *et al.*, 2021; NEMEROW, 1991).

This index considers the mean and maximum values of the Pollution Index (PI) (Equation 6) and highlights chemical species with a high degree of pollution (Equation 7).

$$PI = \frac{C_m}{C_n} \quad (6)$$

$$NPI = \sqrt{\frac{\left(\frac{1}{n} \sum_{i=1}^n PI\right)^2 + PI_{max}^2}{n}} \quad (7)$$

In which:

PI = the calculated single pollution index;

C_m = the metal concentration in the sediment;

C_n = the baseline concentration;

PI_{max} = the maximum value of PI;

n = the number of chemical species considered in the study.

The NPI value is dimensionless and represents the overall pollution provided by a single parameter. As for the degree of pollution, we have that: NPI ≤ 0.5 (sediment no pollution); 0.5 < NPI ≤ 0.7 (sediment clean); 0.7 < NPI ≤ 1.0 (sediment warm); 1.0 < NPI ≤ 2.0 (sediment polluted); 2.0 < NPI ≤ 3.0 (medium pollution); NPI > 3.0 (sediment severe pollution).

RESULTS AND DISCUSSION

Quantifying major oxides in sediment samples

Table 4 shows the average concentrations of major oxides present in the fine fractions (< 63 μm) of the sediment samples measured by WD-XRF.

The statistical results in Table 4 show the variation of the average concentrations of major oxides in the sediment samples considering the spatial variations between the points and the seasonality between the sampling campaigns 1, 2, and 3 (Supplementary Table 1).

The reliability of the results generated with WD-XRF technique (BRUKER, model S8 TIGER) is reinforced by comparing the mean concentrations present in the samples with those presented in the study by Silva *et al.* (2019), who also evaluated the presence of major oxides in sediment samples from the region using the ED-XRF instrumental technique (Shimadzu, model EDX-720).

Table 4 shows that the Na₂O, MnO and SiO₂ concentrations varied significantly between the Barra Funda (upstream the mining zone) sampling point and the others. Also, the MgO and CaO concentrations are significantly different between the Barra Funda (upstream the mining zone) and Rodeio Bonito (beginning of the mining zone) sampling points. Table 4 shows that the P₂O₅ concentration varied significantly between Iraí — BR-386 (IBR) (downstream the mining zone) and the other sampling points.

Statistical analyses of sample datasets

To better explore the WD-XRF analysis results of the fine fractions (< 0.063 mm) of the sediment samples, the sample dataset was subjected to the Grubbs and Jarque-Bera tests (two-tailed, significant p ≥ 0.05). The results show that the datasets are normally distributed and follow the Gaussian model, thus allowing to use parametric statistics (Supplementary Table 2).

Also, the datasets were submitted to ANOVA to identify significant differences between average sediment samples from different sampling points (upstream and downstream of the mining zone) and sampling campaign (seasonal variation), as shown in Table 5.

The results indicated no significant differences for the Fe₂O₃, CuO, SrO, and ZnO major oxides, but significant for the other analytes. The Tukey test (two-tailed, significant p ≥ 0.05) results shown in Table 6 identified the major oxides with significant variance in the datasets.

The results also indicate a significant variance of the major oxides Al₂O₃ and TiO₂ between Iraí — Ponte Velha (IPV) (mining zone) and other sampling points. Figure 4 shows the results for the PCA of the sediment samples (Table 4).

In Figure 4, the PCA of the first two components explains 61.7% of the dataset variance. The negative direction of Principal Component (PC) 1 (39.7%) suggests the differentiation of samples with higher levels of MgO, K₂O, SiO₂, CaO, and TiO₂. The third quadrant clusters the IPV sample point (leaving the mining zone) with high ZnO and TiO₂ concentrations.

The fourth quadrant clusters the samples from the Barra Funda (BF) sampling point with higher levels of Fe₂O₃ and CuO. In general, the PCA plots separated the samples into three distinct clusters, the IPV sampling points (leaving the mining zone), other samples, and BF (upstream the mining zone), according

Table 4 - Average concentrations of major oxides in the Várzea river sediment samples.

	ANALYTE	BF		RB		A		IPV		IBR	
		($\bar{x} \pm CI$)	VC (%)	($\bar{x} \pm CI$)	VC (%)	($\bar{x} \pm CI$)	VC (%)	($\bar{x} \pm CI$)	VC (%)	($\bar{x} \pm CI$)	VC (%)
1ª Sampling campaign	Na_2O	016 ± 001	45	029 ± 001	27	032 ± 001	19	030 ± 001	29	025 ± 006	206
	MgO	049 ± 002	33	067 ± 001	03	066 ± 001	13	060 ± 002	29	060 ± 002	29
	Al_2O_3	162 ± 02	13	139 ± 01	09	139 ± 02	13	125 ± 02	14	156 ± 03	15
	SiO_2	414 ± 06	13	449 ± 02	04	439 ± 03	05	445 ± 11	21	435 ± 05	11
	P_2O_5	027 ± 001	04	025 ± 001	18	023 ± 001	04	026 ± 001	19	030 ± 001	09
	K_2O	037 ± 002	06	055 ± 001	02	056 ± 001	05	056 ± 002	28	053 ± 001	05
	CaO	10 ± 001	09	14 ± 001	07	15 ± 000	03	13 ± 001	09	12 ± 002	19
	TiO_2	65 ± 01	08	91 ± 01	13	77 ± 01	11	97 ± 13	114	45 ± 02	33
	MnO	028 ± 001	07	030 ± 001	07	029 ± 001	05	033 ± 001	17	040 ± 002	43
	Fe_2O_3	177 ± 004	02	168 ± 004	02	170 ± 001	03	178 ± 02	10	164 ± 08	45
	CuO	004 ± 001	12	003 ± 001	24	004 ± 001	18	004 ± 001	21	003 ± 001	48
2ª Sampling campaign	ZnO	002 ± 0001	22	002 ± 0001	18	002 ± 0001	19	004 ± 0001	50	002 ± 0001	27
	SrO	00096 ± 00001	10	001 ± 0001	19	001 ± 0001	29	001 ± 0001	52	001 ± 0001	76
	Na_2O	019 ± 001	37	019 ± 001	58	020 ± 001	30	029 ± 001	19	020 ± 004	17
	MgO	051 ± 002	35	045 ± 001	09	049 ± 002	30	058 ± 001	04	047 ± 001	28
	Al_2O_3	161 ± 02	12	135 ± 01	07	136 ± 02	12	123 ± 01	08	145 ± 02	09
	SiO_2	418 ± 03	06	453 ± 02	04	441 ± 03	07	433 ± 02	05	408 ± 04	10
	P_2O_5	026 ± 001	12	030 ± 001	04	027 ± 001	04	024 ± 001	15	029 ± 001	04
	K_2O	039 ± 001	02	045 ± 001	06	048 ± 001	05	050 ± 001	08	047 ± 001	08
	CaO	107 ± 002	14	102 ± 001	05	106 ± 002	17	118 ± 000	03	100 ± 001	11
	TiO_2	66 ± 01	10	64 ± 01	04	67 ± 02	25	121 ± 05	37	50 ± 01	18
	MnO	025 ± 001	09	037 ± 001	09	035 ± 001	06	032 ± 001	04	038 ± 001	02
3ª Sampling campaign	Fe_2O_3	180 ± 01	04	170 ± 02	12	171 ± 01	05	177 ± 01	04	176 ± 01	04
	CuO	004 ± 001	12	003 ± 001	37	003 ± 001	18	003 ± 001	26	004 ± 001	14
	ZnO	002 ± 001	45	002 ± 001	15	002 ± 001	32	002 ± 001	43	003 ± 001	08
	SrO	001 ± 00001	32	001 ± 000001	15	001 ± 00001	09	000 ± 00001	114	001 ± 00001	29
	Na_2O	016 ± 001	33	030 ± 005	128	020 ± 001	61	027 ± 001	39	020 ± 001	63
	MgO	046 ± 001	06	070 ± 002	27	049 ± 003	60	053 ± 001	21	050 ± 001	05
	Al_2O_3	140 ± 01	09	139 ± 020	11	140 ± 05	19	122 ± 01	10	146 ± 01	02
	SiO_2	427 ± 02	04	446 ± 05	11	416 ± 08	16	444 ± 05	10	431 ± 01	01
	P_2O_5	024 ± 001	04	025 ± 001	13	027 ± 001	24	025 ± 001	08	029 ± 001	13
	K_2O	037 ± 001	08	060 ± 001	03	047 ± 001	05	055 ± 002	31	049 ± 001	04
	CaO	107 ± 001	03	155 ± 002	09	113 ± 002	12	125 ± 001	10	106 ± 001	04
TiO_2	700 ± 03	34	785 ± 003	06	650 ± 003	08	850 ± 06	112	700 ± 002	06	
MnO	028 ± 001	10	030 ± 001	01	031 ± 001	04	032 ± 001	14	034 ± 001	07	
Fe_2O_3	166 ± 01	05	170 ± 01	03	170 ± 01	04	169 ± 01	05	174 ± 02	09	
CuO	004 ± 001	12	003 ± 001	12	003 ± 001	18	003 ± 001	10	003 ± 001	40	
ZnO	002 ± 001	30	002 ± 001	18	002 ± 001	24	002 ± 001	24	002 ± 001	52	
SrO	001 ± 0001	33	001 ± 0001	20	001 ± 0001	29	001 ± 0001	103	001 ± 0001	52	

\bar{x} : average concentrations; CI: confidence interval; VC (%): variance coefficient; BF: Barra Funda; RB: Rodeio Bonito; A: Ametista do Sul; IPV: Irai – Ponte Velha; IBR: Irai – BR-386. A total of 45 WD-XRF analyses were performed on the fine fractions of 15 composite sediment samples from the Várzea river.

Source: elaborated by the authors.

Table 5 – Statistical variance analysis by Analysis of Variance (two-tailed, significant $p \geq 0.05$).

Analyte	Variation source	SQ	DF	MQ	F	p value
Na ₂ O	<i>Between groups</i>	0.1	4	0.0	8.9	3.09 x 10 ⁻⁵
	<i>Within groups</i>	0.1	40	0.0		
	<i>Total</i>	0.1	44	0.0		
MgO	<i>Between groups</i>	0.1	4	0.0	3.3	0.02
	<i>Within groups</i>	0.2	40	0.0		
	<i>Total</i>	0.3	44	0.0		
Al ₂ O ₃	<i>Between groups</i>	53.8	4	13.4	43.6	4.5 x 10 ⁻¹⁴
	<i>Within groups</i>	12.3	40	0.3		
	<i>Total</i>	66.2	44	0.0		
SiO ₂	<i>Between groups</i>	52.3	4	13.1	15.0	1.45 x 10 ⁻⁷
	<i>Within groups</i>	34.9	40	0.9		
	<i>Total</i>	87.2	44	0.0		
P ₂ O ₅	<i>Between groups</i>	0.0	4	0.0	9.7	1.40 x 10 ⁻⁵
	<i>Within groups</i>	0.0	40	0.0		
	<i>Total</i>	0.0	44	0.0		
K ₂ O	<i>Between groups</i>	0.1	4	0.0	23.7	4.12 x 10 ⁻¹⁰
	<i>Within groups</i>	0.1	40	0.0		
	<i>Total</i>	0.2	44	0.0		
CaO	<i>Between groups</i>	0.5	4	0.1	6.2	0.00
	<i>Within groups</i>	0.8	40	0.0		
	<i>Total</i>	1.3	44	0.0		
TiO ₂	<i>Between groups</i>	101.9	4	25.5	20.1	3.75 x 10 ⁻⁹
	<i>Within groups</i>	50.6	40	1.3		
	<i>Total</i>	152.5	44	0.0		
MnO	<i>Between groups</i>	0.0	4	0.0	23.1	6.09 x 10 ⁻¹⁰
	<i>Within groups</i>	0.0	40	0.0		
	<i>Total</i>	0.1	44	0.0		
Fe ₂ O ₃	<i>Between groups</i>	2.15	4	0.5	2.5	0.05
	<i>Within groups</i>	8.5	40	0.2		
	<i>Total</i>	10.7	44	0.6		

SQ: sum of squares; DF: degrees of freedom; MQ: mean square; F: Fisher value.
Source: elaborated by the authors.

to the variables (chemical compounds) that contributed to this grouping as shown in the score plots.

The hierarchical cluster analysis generated a dendrogram (Figure 5) similar to the two large clusterings generated by the PCA analysis.

The two different clusters generated by similarity (agglomeration coefficient 0.8) suggest similar geochemical characteristics of these samples. The dendrogram grouped the sediment samples from different sampling campaigns into two clusters, the BF sampling point (upstream the mining zone) and the other sampling points while, by similarity, it grouped the current concentrations of the IPV sampling point (leaving the mining zone).

Determining the environmental geochemical indexes

Table 7 shows the environmental geochemical indices CIA, ER1, and ER2, calculated based on the average concentrations of major oxides. The CIA scores calculated for the Várzea river sediment samples demonstrate extreme chemical

weathering, a condition typical of hot and humid regions (GOMES; ALMEIDA; SPERANDIO, 2018; KUSIN *et al.*, 2019).

These results were expected since the matrix analysis characteristics indicated a significant difference between the fluvial sediments and the soil of the watershed due to the heterogeneous mixture of grains caused by the significant transport that weathered particles undergo as a result of the agents such as water (AMOSOVA *et al.*, 2019; SHEN *et al.*, 2018; SILVA *et al.*, 2019).

Moreover, according to the Köppen classification, the study region climate is subtropical with hot summers, without prolonged droughts, and average rainfall distributed throughout the year (SEMA, 2012; SILVA *et al.*, 2019), thus favoring the weathering processes. The scores associated with RE₁ and RE₂ elementary ratios allow identifying the source rock composition (SANG *et al.*, 2018; SOUSA *et al.*, 2018).

Thus, the ER₁ values associated with sediment samples from all sampling points of the Várzea river suggest origin associated with clay minerals. Also, the

Table 6 - Tukey test (two-tailed, significant p 0.05) of major oxides in the sediment samples.

Na₂O	BF	RB	A	IPV	IBR	P₂O₅	BF	RB	A	IPV	IBR
BF		0.0	0.01	0.0	0.1	BF		0.5	1.0	1.0	0.0
RB	6.5		0.7	0.9	0.3	RB	2.3		0.8	0.2	0.0
A	4.8	1.7		0.2	0.9	A	0.8	1.4		0.8	0.0
IPV	7.7	1.2	2.9		0.0	IPV	0.8	3.1	1.6		0.0
IBR	3.7	2.8	1.1	4.0		IBR	7.0	4.8	6.2	7.9	
MgO	BF	RB	A	IPV	IBR	K₂O	BF	RB	A	IPV	IBR
BF		0.0	0.5	0.1	0.7	BF		0.0	0.0	0.0	0.0
RB	4.8		0.4	0.8	0.2	RB	11.7		0.5	1.0	0.3
A	2.3	2.5		0.9	1.0	A	9.4	2.3		0.4	1.0
IPV	3.4	1.4	1.0		0.8	IPV	11.8	0.1	2.4		0.3
IBR	1.7	3.1	0.6	1.7		IBR	8.9	2.8	0.5	2.9	
Al₂O₃	BF	RB	A	IPV	IBR	CaO	BF	RB	A	IPV	IBR
BF		0.0	0.0	0.0	0.2	BF		0.0	0.1	0.1	1.0
RB	9.1		1.0	0.0	0.0	RB	5.9		0.4	0.6	0.0
A	9.1	0		0.0	0.0	A	3.5	2.5		1.0	0.2
IPV	17.2	8.0	8.0		0.0	IPV	3.8	2.1	0.4		0.1
IBR	3.1	6.0	6.0	14.1		IBR	0.4	5.5	3.1	3.5	
SiO₂	BF	RB	A	IPV	IBR	TiO₂	BF	RB	A	IPV	IBR
BF		0.0	0.0	0.0	0.8	BF		0.3	1.00	0.0	0.2
RB	9.6		0.0	0.3	0.0	RB	2.9		0.5	0.0	0.0
A	4.0	5.5		0.3	0.4	A	0.7	2.2		0.0	0.1
IPV	6.8	2.7	2.8		0.0	IPV	8.9	6.0	8.2		0.0
IBR	1.6	8.0	2.5	5.2		IBR	3.2	6.1	3.9	12.1	
MnO	BF	RB	A	IPV	IBR	Fe₂O₃	BF	RB	A	IPV	IBR
BF		0.0	0.0	0.0	0.0	BF		0.1	0.3	1.0	0.7
RB	6.5		1.0	1.0	0.0	RB	3.3		1.0	0.1	0.8
A	5.6	0.8		0.8	0.0	A	2.8	0.5		0.2	0.9
IPV	7.2	0.7	1.5		0.0	IPV	0.2	3.5	3.0		0.6
IBR	13.5	7.0	7.9	6.3		IBR	1.8	1.5	1.0	1.9	

BF: Barra Funda; RB: Rodeio Bonito; A: Ametista do Sul; IPV: Iraí – Ponte Velha; IBR: Iraí – BR-386.
Source: elaborated by the authors.

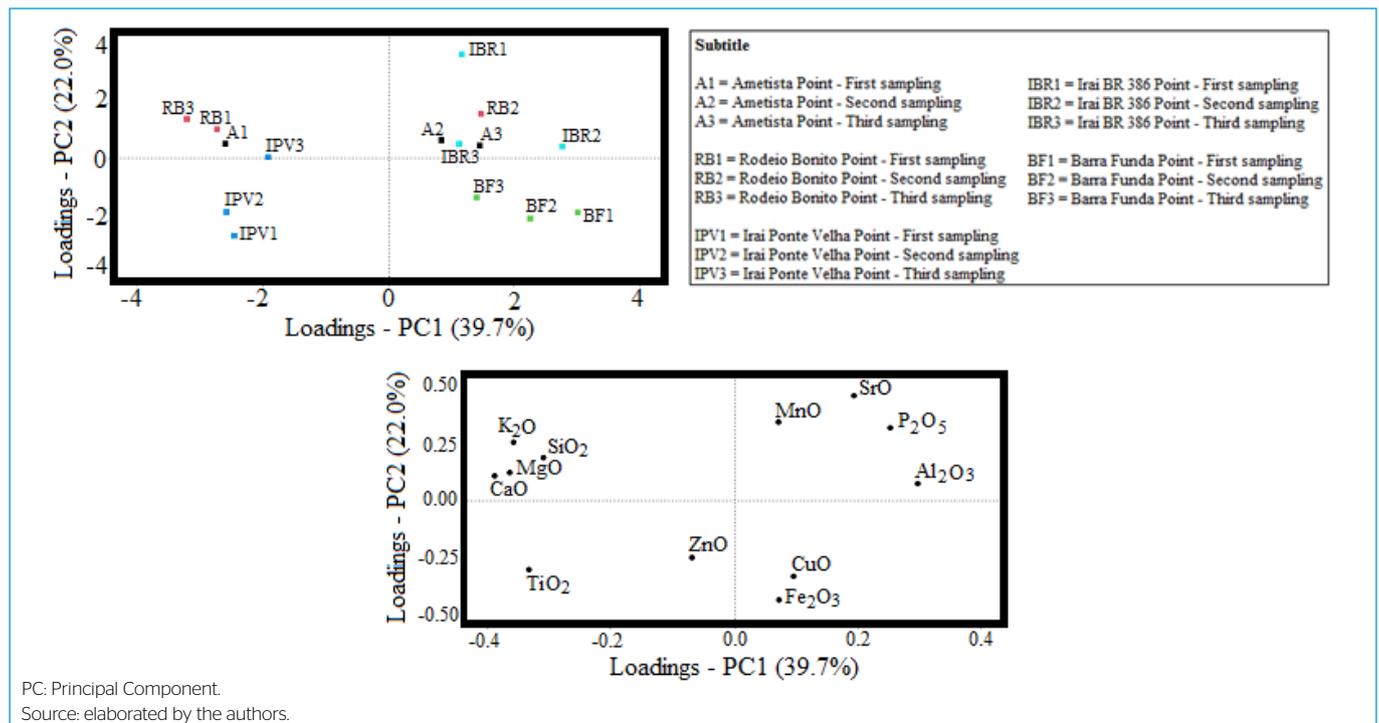


Figure 4 - Principal Component Analysis of the average concentrations of major oxides present in sediment samples from the Várzea river, in the state of Rio Grande do Sul.

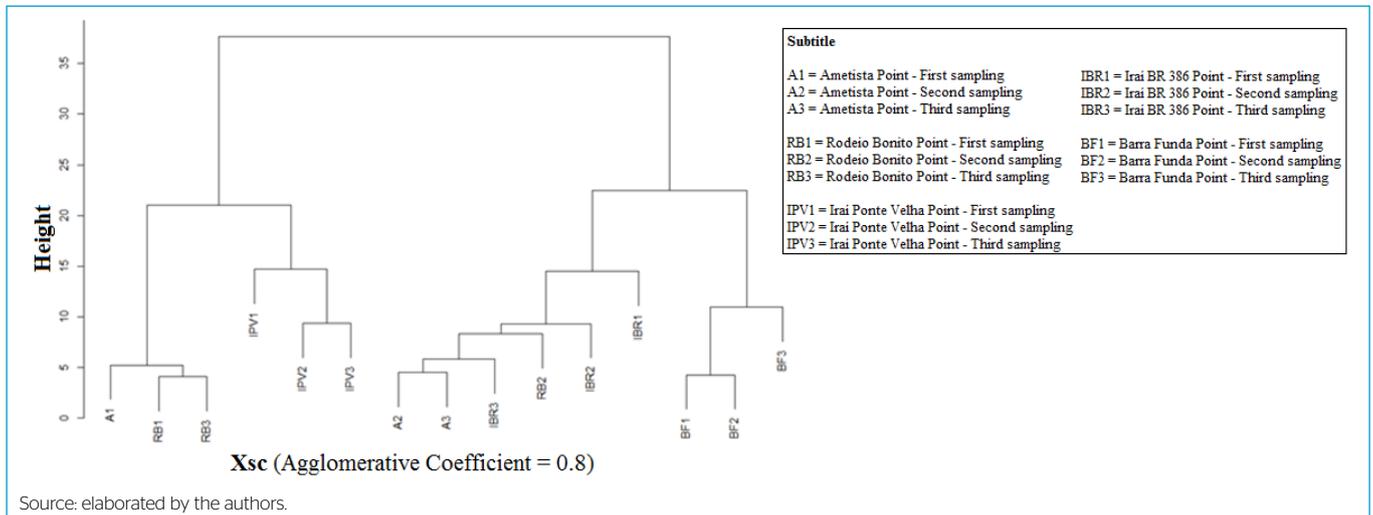


Figure 5 – Dendrogram resulting from the cluster analysis grouping the average concentrations of major oxides present in sediment samples from the Várzea river, in the state of Rio Grande do Sul.

Table 7 – Environmental geochemical indexes determined for the Várzea river sediments and basalt rocks from tailings in the mining region.

	Chemical Index of Alteration (CIA)	Elementary Ratio (ER ₁)	Elementary Ratio (ER ₂)
Sediments			
Barra Funda (BF)	90.6	0.03	2.3
Rodeio Bonito (RB)	86.6	0.04	1.8
A (Ametista)	87.5	0.04	2.0
IPV (Irai – Ponte Velha)	85.6	0.04	1.2
IBR (Irai – BR-386)	89.2	0.03	2.7
Mining tailings			
Baggio <i>et al.</i> (2015)	50.0	0.09	5.7
Hartmann <i>et al.</i> (2015)	56.6	0.01	1.2
Korchagin; Caner; Bortoluzzi (2019)	54.2	0.03	4.5

Source: elaborated by the authors.

ER₂ values for sediments from all sampling sites indicate a mafic igneous rock origin such as basalt rocks, for example (GOMES; ALMEIDA; SPERANDIO, 2018; KUSIN *et al.*, 2019; SANG *et al.*, 2018).

The ER₁ and ER₂ values shown in Table 7 suggest an origin associated with clay minerals and mafic igneous rocks, respectively. Likewise, HARTMANN *et al.* (2015) reported similar values for the sediment samples. ER₃ (with values ranging from 2.7 to 3.6) suggests a low compositional maturity of the sediments in the study region. The ER₄ values, ranging between 31.8 and 46.6, indicate the existence of variations in the chemical composition between the sampling points of the Várzea river, in the fine fractions (< 0.063 mm) of the sediment samples. The total content of CaO is different from the other study regions at the Ametista do Sul sampling point (A), mining zone (ARISEKAR *et al.*, 2021; BESSA *et al.*, 2020; SILVA; BATEZELLI; LADEIRA, 2015; ZANARDO *et al.*, 2017). These results are shown in Table 8.

Figure 6 presents the K₂O/Na₂O versus SiO₂ graph that defines the potential geochemical similarity of sediments in the study region, with which is possible to generate a behavioral distinction established through the relationship between the concentrations of major oxides that constitute the fine fractions of the sediment samples.

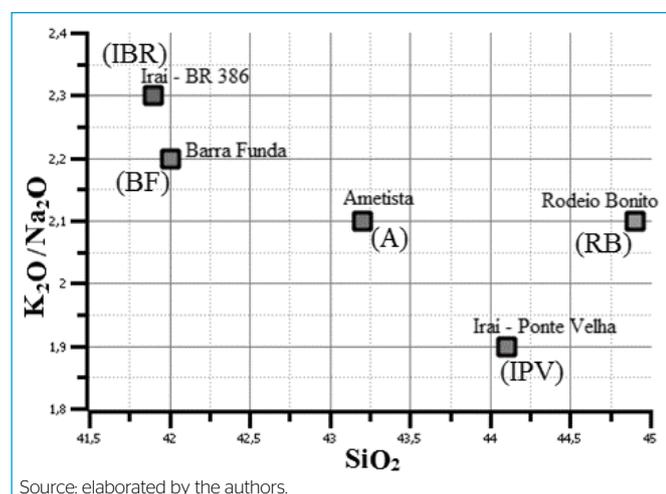
Table 8 – Environmental geochemical indexes and CaO content determined for the Várzea river sediments and basalt rocks from tailings in the mining region.

	Elementary ratio (ER ₃)	Elementary ratio (ER ₄)	CaO
Sediments			
Barra Funda (BF)	2.7	46.3	1.0
Rodeio Bonito (RB)	3.3	31.8	1.3
A (Ametista)	31	33.8	1.5
IPV (Irai – Ponte Velha)	3.6	32.5	1.2
IBR (Irai – BR-386)	2.8	34.5	1.1

Source: elaborated by the authors.

The sampling point of IPV is located immediately after the mineral extraction zone in the municipality of Ametista do Sul, Southern Brazil. The point IPV presents a different behavior, in the graph in Figure 6, from the other sampling points in the water system.

The reference values calculated by the NPI (Table 9) reveal a potential anthropic influence on the water system due to mineral extraction.



Source: elaborated by the authors.

Figure 6 – Geochemical similarity of sediments sampling points from Várzea river, Southern Brazil.

Table 9 – Nemerow Pollution Index – environmental pollution of sediment samples.

Major oxides	Nemerow Pollution Index				
	BF	RB	A	IPV	IBR
Al_2O_3	1.2	1.0	1.0	0.9	1.0
SiO_2	1.5	1.6	1.6	1.6	1.2
P_2O_5	0.9	1.0	0.9	0.9	1.0
K_2O	1.2	1.7	1.6	1.7	1.2
CaO	1.1	1.5	1.3	1.3	1.1
TiO_2	1.1	1.3	1.1	1.7	1.0
MnO	0.8	1.0	1.0	1.0	1.1
Fe_2O_3	1.1	1.0	1.0	1.1	1.0
CuO	3.0	2.5	2.7	2.9	1.5
ZnO	2.5	2.5	2.4	2.5	1.3

BF: Barra Funda; RB: Rodeio Bonito; A: Ametista do Sul; IPV: Iraí – Ponte Velha; IBR: Iraí – BR-386.

Source: elaborated by the authors.

The baseline (background values) used in this study for the region were determined by Silva *et al.* (2019). The mining tailings around the Várzea river extraction region are basalt rocks with dark tones and are rich in iron, magnesium, titanium, and silica oxides (BAGGIO *et al.*, 2015; KORCHANGIN *et al.*, 2019).

The NPI values can be associated with pluviometric carry processes of the major oxides SiO_2 , K_2O , CuO , ZnO present in the sediments embedded mining tailings, indicating the mobility of the content of the basalt rock deposits (ARISEKAR *et al.*, 2021; BESSA *et al.*, 2020). The calculated NPI values show that the Várzea river sediment samples range from polluted for the major oxides CuO and ZnO (point of Iraí BR-386 – IBR), TiO_2 (point of IPV), K_2O (points of IPV, A and Rodeio Bonito – RB) and medium polluted for the major oxides CuO and ZnO (points of Barra Funda – BF, RB, A, and IPV). Furthermore, the NPI values together with the WD-XRF results of the sediment fine fractions from the two sampling campaigns evidence the potential contribution of the SiO_2 major oxides

(upstream the BF sampling point that decreases in the downstream sampling points), K_2O , and CaO (in BF, upstream the mining zone, and at the IPV sampling point, in the mining zone) and TiO_2 (IPV, in the mining zone). In Table 9, the differentiation of NPI values for mineral oxide TiO_2 , CuO and ZnO at the sampling point (IPV) is highlighted, exiting the mine extraction zone, indicating the pollution of the sediment samples. These concentrations present in sediment samples are often used to identify the provenance and source tectonic environment, as they are relatively immobile during the sedimentary process.

The K_2O is a nutrient that is abundant in the earth's crust and is resistant to weathering processes. Already the CuO is present in range of silicates present in the drainage sediments. In mineral extraction regions, rich in hydrated Fe and Mn oxides, CuO is removed from the drainage sediments by adsorption and recitation in response to changes in pH and Eh. ZnO mobilization is limited during the periods of weathering changes in volcanic rocks and losses due to metamorphism (BAGGIO *et al.*, 2015; BRANCO, 2002; HARTMANN *et al.*, 2015; MINEROPAR, 2001).

Samples of sedimentary rocks and river sediments are usually predominantly characterized by much higher average concentrations of SiO_2 and Al_2O_3 compared to those of K_2O and TiO_2 (BECK *et al.*, 2020; KUSIN *et al.*, 2019; SANG *et al.*, 2018). However, the average concentrations of major oxides determined in the mining zone region of the Várzea river indicate otherwise, thus reinforcing the possibility of anthropic input due to mining activities from semi-precious stones.

CONCLUSIONS

The average concentrations of major oxides such as Al_2O_3 , MnO , CaO , SiO_2 , K_2O , and TiO_2 present in the fractions of sediment samples from the Várzea river, in Southern Brazil, varied significantly ($p < 0.05$) between the studied sampling points of A, in the mining zone, and IPV, leaving the mining zone. The PCA and the cluster analysis also suggest the existence of three distinct mineral oxide groups, differentiating IPV (leaving the mining zone) and BF (upstream the mining zone) from the other sampling points.

The CIA results indicate extreme chemical weathering of the fluvial sediment samples from the Várzea river. The ER_1 and ER_2 elementary ratios suggest that clay minerals and mafic igneous rocks formed the sediment samples. Also, there is an evident similarity between the ER_2 values of the sediments and those of the basalt rock tailings from the mining zone.

The ER_3 suggests a low compositional maturity, the ER_4 indicate the existence of variations in the chemical composition of the sediments. The content of CaO is different at the sampling point A, mining zone. The K_2O/Na_2O versus SiO_2 graph showed the distinction between the concentrations of major oxides of the sediment samples in IPV.

Regarding, the NPI of the sediments suggests that SiO_2 , K_2O , CaO , CuO , ZnO and TiO_2 major oxides occur especially at the IPV sampling point, classifying the sediment samples as from polluted to medium polluted. The results suggest that the environmental stress caused by major oxides supplied to the water system originates from basaltic rock tailings accumulated in front of the mines (sediment formation areas), the byproduct of the mining processes.

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AUTHORS' CONTRIBUTIONS

Silva, P.R.B.: Conceptualization, Data curation, Investigation, Methodology, Validation, Writing – original draft. Parizotto, D.: Writing – original draft, Writing – review & editing. Silva, L.R.: Writing – review & editing. Parreira, P.S.: Data curation, Formal analysis. Melquiades, F.L.: Data curation, Formal analysis. Mauad, F.F.: Project administration, Supervision.

REFERENCES

- ALVES, W.; FERREIRA, P.; ARAÚJO, M., Challenges and pathways for Brazilian mining sustainability. *Resources Policy*, v. 74, 101648, 2021. <https://doi.org/10.1016/j.resourpol.2020.101648>.
- AMOSOVA, A.A.; CHUBAROV, V.M.; PASHKOVA, G.V.; FINKELSHTEIN, A.L.; BEZRUKOVA, E.V. Wavelength dispersive X-ray fluorescence determination of major oxides in bottom and peat sediments for paleoclimatic studies. *Applied Radiation and Isotopes*, v. 144, p. 118-123, 2019. <https://doi.org/10.1016/j.apradiso.2018.11.004>
- ARISEKAR, U.; SHAKILA, R.J.; SHALINI, R.; SIVARAMAN, B.; JEYASEKARAN, G.; MALINI, N.A.H. Heavy metal concentration in reef-associated surface sediments, Hare Island, Gulf of Mannar Marine Biosphere Reserve (southeast coast of India): The first report on pollution load and biological hazard assessment using geochemical normalization factors and hazard indices. *Marine Pollution Bulletin*, v. 162, 111838, 2021. <https://doi.org/10.1016/j.marpolbul.2020.111838>
- BAGGIO, S.B.; HARTMANN, L.A.; MASSONNE, H.J.; THEVE, T.; ANTUNES, L.M. Silica gossan as a prospective guide for amethyst geode deposits in the Ametista do Sul mining district, Paraná volcanic province, southern Brazil. *Journal of Geochemical Exploration*, v. 159, p. 213-226, 2015. <https://doi.org/10.1016/j.gexplo.2015.09.011>
- BECK, K.K.; MARIANI, M.; FLETCHER, M.S.; SCHNEIDER, L.; AQUINO-LÓPEZ, M.A.; GADD, P.S.; HEIJNIS, H.; SAUNDERS, K.M.; ZAWADZKI, A. The impacts of intensive mining on terrestrial and aquatic ecosystems: A case of sediment pollution and calcium decline in cool temperate Tasmania, Australia. *Environmental Pollution*, v. 265, 114695, 2020. <https://doi.org/10.1016/j.envpol.2020.114695>
- BESSA, A.Z.E.; NGUETCHOUA, G.; JANPOU, A.K.; EL-AMIER, Y.A.; NGUETCHOUA, O.A.N.N.M.; KAYOU, U.R.K.; BISSE, S.B.; MAPUNA, E.C.N.; AMSTRONG-ALTRIN, J.S. Heavy metal contamination and its ecological risks in the beach sediments along the Atlantic Ocean (Limbe coastal fringes, Cameroon). *Earth Systems and Environment*, v. 5, p. 433-444, 2020. <https://doi.org/10.1007/s41748-020-00167-5>
- BILLAH, M.M.; KOKUSHI, E.; UNO, S. Distribution, geochemical speciation, and bioavailable potencies of cadmium, copper, lead, and zinc in sediments from urban coastal environment in Osaka Bay, Japan. *Water, Air, & Soil Pollution*, v. 230, p. 1-13, 2019. <https://doi.org/10.1007/s11270-019-4196-8>
- BRANCO, P.M. *Mapa geológico do estado do Rio Grande do Sul*. Porto Alegre: CRPM Serviço Geológico do Brasil, 2002. 45 p.
- BRANDÃO, C.J.; BOTELHO, M.J.C.; SATO, M.I.Z.; LAMPARELLI, M.C. (org.). *Guia nacional da coleta e preservação de amostras: água, sedimento, comunidades aquáticas e efluentes líquidas*. São Paulo: CETESB; Brasília: ANA, 2011.
- BURTON, G.A.; JOHNSTON, E. Avaliação de sedimentos contaminados no contexto de Múltiplos Estressores. *Environmental Toxicology and Chemistry*, v. 29, n. 12, p. 2625-2646, 2010.
- CROUDACE, I.W.; LÖWEMARK, L.; TJALLINGII, R.; ZOLITSCHKA, B. Current perspectives on the capabilities of high resolution XRF core scanners. *Quaternary International*, v. 514, p. 5-15, 2019. <https://doi.org/10.1016/j.quaint.2019.04.002>
- DIAS, C.H.; CHAVES, M.L.S.C.; JUCHEM, P.L.; ROMANO, A.W. Ocorrências de ametista em basaltos do Triângulo Mineiro (Minas Gerais): descrição e comparações com depósitos similares do Rio Grande do Sul. *Pesquisas em Geociências*, v. 46, n. 3, p. 1-19, 2019. <https://doi.org/10.22456/1807-9806.97385>
- DOTMATIC. *GraphPad Software: Quick Calcs Outlier Calculator*. Dotmatics, 2020. Available at: <https://graphpad.com/quickcalcs/Grubbs1.cfm>. Accessed on: Jan. 8, 2020.
- FUNDAÇÃO ESTADUAL DE PROTEÇÃO AMBIENTAL HENRIQUE LUIS ROESSLER (FEPAM). *Bacia hidrográfica do Rio da Várzea*. Fepam, 2020. Available at: https://www.fepam.rs.gov.br/qualidade/bacia_uru_varzea.asp. Accessed on: Jan. 8, 2020.
- GLOAGUEN, T.V.; MOTTA, P.N.S.D.; COUTO, C.F. A grain-size correction for metal pollution indexes in river sediments. *International Journal of Sediment Research*, v. 36, n. 3, p. 362-372, 2021. <https://doi.org/10.1016/j.ijsrc.2020.10.005>
- GOMES, C.H.; ALMEIDA, D.P.M.; SPERANDIO, D.G. Geoquímica de Sedimentos da Confluência das Bacias Hidrográficas Baixo Jacuí e Vacacaí-Mirim, Caçapava do Sul-RS: Implicações para Proveniência e Intemperismo Químico. *Anuário do Instituto de Geociências*, v. 41, n. 3, p. 470-482, 2018. https://doi.org/10.11137/2018_3_470_482
- GUILHEN, S.N.; COTRIM, M.E.B.; SAKATA, S.K.; SCAPIN, M.A. Application of the fundamental parameter method to the assessment of major and trace elements in soil and sediments from Osamu Utsumi uranium mine by WDXRF. *REM: International Engineering Journal*, v. 72, n. 4, p. 609-617, 2019. <https://doi.org/10.1590/0370-44672018720146>
- HAMMER, Ø. *Past paleontological statistics*. Version 3.08. Oslo: Natural History Museum University of Oslo, 2015.
- HARTMANN, L.A.; MEDEIROS, J.T.N.; BAGGIO, S.B.; ANTUNES, L.M. Controls on prolate and oblate geode geometries in the Veia Alta basalt flow, largest world producer of amethyst, Paraná volcanic province, Brazil. *Ore Geology Reviews*, v. 66, p. 243-251, 2015. <https://doi.org/10.1016/j.joregeorev.2014.11.005>
- HARTMANN, L.A.; PERTILLE, J.; DUARTE, L.C. Giant-geode endowment of tumuli in the Veia Alta flow, Ametista do Sul. *Journal of South American Earth Sciences*, v. 77, p. 51-57, 2017. <https://doi.org/10.1016/j.jsames.2017.04.013>
- HERATH, D.; PITAWALA, A.; GUNATILAKE, J.; IQBAL, M.C.M. Using multiple methods to assess heavy metal pollution in an urban city. *Environmental Monitoring AND Assessment*, v. 190, p. 657-671, 2018. <https://doi.org/10.1007/s10661-018-7016-5>

- INSTITUTO NACIONAL DE METROLOGIA, QUALIDADE E TECNOLOGIA (INMETRO). *Orientação sobre validação de métodos analíticos*: documento de caráter orientativo. DOQ-CGCRE-008. Revisão 5 - Agosto de 2016. Rio de Janeiro: INMETRO, 2018.
- KORCHAGIN, J.; CANER, L.; BORTOLUZZI, E.C. Variability of amethyst mining waste: A mineralogical and geochemical approach to evaluate the potential use in agriculture. *Journal of Cleaner Production*, v. 210, p. 749-758, 2019. <https://doi.org/10.1016/j.jclepro.2018.11.039>
- KUSIN, F.M.; AWANG, N.H.C.; HASAN, S.N.M. S.; RAHIM, H.A.A.; AZMIN, N.; JUSOP, S.; KIM, K.W. Geo-ecological evaluation of mineral, major and trace elemental composition in waste rocks, soils and sediments of a gold mining area and potential associated risks. *CATENA*, v. 183, 104229, 2019. <https://doi.org/10.1016/j.catena.2019.104229>
- LONE, A.M.; ACHYUTHAN, H.; SHAH, R.A.; SANGODE, S.J. Environmental magnetism and heavy metal assemblages in lake bottom sediments, Anchar Lake, Srinagar, NW Himalaya, India. *International Journal of Environmental Research*, v. 12, p. 489-502, 2018. <https://doi.org/10.1007/s41742-018-0108-9>
- MINERAIS DO PARANÁ S.A. (MINEROPAR). *Atlas Geoquímico da Folha de Curitiba*. Curitiba: Folha de Curitiba, 2001. 80 p.
- MIZAEI, J.O.S.S.; CARDOSO-SILVA, S.; FRASCARELI, D.; POMPÉO, M.L.M.; MOSCHINI-CARLOS, V. Ecosystem history of a tropical reservoir revealed by metals, nutrients and photosynthetic pigments preserved in sediments. *CATENA*, v. 184, 104242, 2020. <https://doi.org/10.1016/j.catena.2019.104242>
- MUSEU DE SOLOS DO RIO GRANDE DO SUL (MSRS). *Solos de Planalto*: unidade de Passo Fundo. 2020. Available at: <https://www.ufsm.br/museus/msrs>. Accessed on: Jan. 10, 2020.
- NEMEROW, N.L. *Stream, Lake, Estuary and Ocean Pollution*. 2 ed. New York: Van Nostrand Reinhold, 1991. 472 p.
- OYEDOTUN, T.D.T. X-ray fluorescence (XRF) in the investigation of the composition of earth materials: a review and an overview. *Geology, Ecology and Landscapes*, v. 2, n. 2, p. 17, 2018. <https://doi.org/10.1080/24749508.2018.1452459>
- PASTRAN, S.H.; MALLETT, A. Unearthing power: A decolonial analysis of the Samarco mine disaster and the Brazilian mining industry. *The Extractive Industries and Society*, v. 7, n. 2, p. 704-715, 2020. <https://doi.org/10.1016/j.exis.2020.03.007>
- REMOR, M.B.; SAMPAIO, S.C.; RIJK, S.; BOAS, M.A.V.; GOTARDO, J.T.; PINTO, E.T.; SHARDONG, F.A. Sediment geochemistry of the urban Lake Paulo Gorski. *International Journal of Sediment Research*, v. 33, n. 4, p. 406-414, 2018. <https://doi.org/10.1016/j.ijsrc.2018.04.009>
- SANG, P.N.; LIU, Z.; ZHAO, Y.; ZHAO, X.; PHA, P.D.; VAN LONG, H. Chemical weathering in central Vietnam from clay mineralogy and major-element geochemistry of sedimentary rocks and river sediments. *Heliyon*, v. 4, n. 7, e00710, 2018. <https://doi.org/10.1016/j.heliyon.2018.e00710>
- SECRETARIA DO AMBIENTE E DO DESENVOLVIMENTO SUSTENTÁVEL (SEMA). *Relatório anual sobre a situação dos recursos hídricos no Estado do Rio Grande do Sul*. Porto Alegre: Governo do Estado do Rio Grande do Sul, 2012.
- SANTOS, H.G.; JACOMINE, P.K.T.; ANJOS, L.H.C.; OLIVEIRA, V.A.; LUMBRERAS, J.F.; COELHO, M.R.; ALMEIDA, J.A.; ARAÚJO FILHO, J.C.; OLIVEIRA, J.B.; CUNHA, T.J.F. *Sistema Brasileiro de Classificação de Solos*. 5 ed. rev. e ampl. Brasília: EMBRAPA, 2018. 356 p.
- SHEN, M.; GUO, H.; JIA, Y.; CAO, Y.; ZHANG, D. Partitioning and reactivity of iron oxide minerals in aquifer sediments hosting high arsenic groundwater from the Hetao basin, P. R. China. *Applied Geochemistry*, v. 89, p. 190-201, 2018. <https://doi.org/10.1016/j.apgeochem.2017.12.008>
- SILVA, M.L.; BATEZELLI, A.; LADEIRA, F.S.B. Índices de intemperismo e evolução dos paleossolos da Formação Marília, Maastrichtiano da Bacia neocretácea Bauru. *Geochimica Brasiliensis*, v. 29, n. 2, p. 127-138, 2015.
- SILVA, P.R.B.; MAKARA, C.N.; MUNARO, A.P.; SCHNITZLER, D.C.; WASTOWSKI, A.D.; POLETO, C. Comparison of the analytical performance of EDXRF and FAAS techniques in the determination of metal species concentrations using protocol 3050B (USEPA). *International Journal of River Basin Management*, v. 14, n.4, p. 401-406, 2016. <https://doi.org/10.1080/15715124.2016.1203792>
- SILVA, P.R.B.; DALLA NORA, F.E.; CASTRO, R.J.; WASTOWSKII, A.D.; MAUAD, F.F. The environmental quality of sediments of rivers near prospection areas of semiprecious rocks. *Environmental Monitoring and Assessment*, v. 191, 364, 2019. <https://doi.org/10.1007/s10661-019-7456-6>
- SOURCEFORGE. *Software de geração gráfica SciDAVIS*. 2018. Available at: <https://scidavis.sourceforge.net>. Accessed on: Jan. 2, 2020.
- SOUSA, S.S.C.G.; GUEDES, E.; CASTRO, J.W.A.; ÁVILLA, C.A.; NEUMANN, R. Química Mineral de Granada no Estudo da Proveniência Sedimentar do Litoral Norte do Estado do Rio de Janeiro, Brasil. *Anuário do Instituto de Geociências*, v. 41, n. 2, p. 318-331, 2018. https://doi.org/10.11137/2018_2_318_331
- SUNDARARAJAN, S.; KHADANGA, M.K.; KUMAR, J.P.P.J.; RAGHUMARAN, S.; VIJAYA, R.; JENA, B. K. Ecological risk assessment of trace metal accumulation in sediments of Veraval Harbor, Gujarat, Arabian Sea. *Marine Pollution Bulletin*, v. 114, n. 1, p. 592-601, 2017. <https://doi.org/10.1016/j.marpolbul.2016.09.016>
- THE R FOUNDATION. *R - The R Project for Statistical Computing*. [s.d.]. Available at: <https://www.r-project.org>. Accessed on: March 3, 2022
- UNITED STATES ENVIRONMENTAL PROTECTION AGENCY (USEPA). *Framework for ecological risk assessment*. EPA/630/R-92/001. Washington: USEPA, 1992. 58 p.
- VOSOOGH, A.; SAEEDI, M.; LAK, R. Heavy metals relationship with water and size-fractionated sediments in rivers using canonical correlation analysis (CCA) case study, rivers of southwestern Caspian Sea. *Environmental Monitoring and Assessment*, v. 188, 603, 2016. <https://doi.org/10.1007/s10661-016-5611-x>
- ZANARDO, A.; NAVARRO, G.R.B.; MONTIBELLER, C.C.; ROCHA, R.R.; MORENO, M.M.T.; DEL ROVERI, C.; AZZI, A.A. Geoquímica e proveniência dos sedimentos da formação Corumbataí na região de Rio Claro/SP. *Geociências*, v. 36, n. 1, p. 30-47, 2017. <https://doi.org/10.5016/geociencias.v36i112291>
- ZHAO, X.; HE, B.; WU, H.; ZHENG, G.; MA, X.; LIANG, J.; LI, P.; FAN, Q. A comprehensive investigation of hazardous elements contamination in mining and smelting-impacted soils and sediments. *Ecotoxicology and Environmental Safety*, v. 192, 110320, 2020. <https://doi.org/10.1016/j.ecoenv.2020.110320>